



Revisión [Review]

ECOSYSTEM SERVICES IN BRAZILIAN'S SOUTHERN
AGROFORESTRY SYSTEMS¹[SERVICIOS ECOSISTÉMICOS EN SISTEMAS AGROFORESTALES
DEL SUR DE BRASIL]Geraldo Ceni Coelho^{1*}

¹ Universidade Federal da Fronteira Sul - Chapecó,
Av. Fernando Machado, 108 E, 89802-112 Chapecó, Santa Catarina,
Brazil. +55(49) 2049 6528, Email: cenicoelho@gmail.com

*Corresponding author

SUMMARY

Agroforestry systems (AFS) are polycultures with at least one tree species. These systems provide various ecosystem services, income increments and food safety. These ecosystem services include biodiversity conservation, carbon sequestration, reduction of crop diseases, increased biological controls, biological nitrogen fixation and nutrient cycling. A review of potential ecosystem services of AFS on Southern Brazil is presented. In addition, the potential of carbon uptake through conversion to AFS is estimated. The predominant AFS are agroforestry with yerba mate (*Ilex paraguariensis* A. St. Hil.), silvopastures, citrus and banana orchards, and the palm açai-juçara (*Euterpe edulis* Mart.). Considering the conversion of conventional systems to AFS, the silvopastures present the greatest carbon sequestration potentiality due to the great area used for cattle ranching. The conversion of citrus and banana cropping also present great carbon uptake potential besides reducing fungal and bacterial diseases. Southern Brazil presents more than 15 million hectares which could be converted into silvopasture and other AFS by taking as a model the already well-established experiences. Moreover, AFS has become a strategy for forest restoration. There are no negative trade-offs related to the silvopasture and citrus agroforestry adoption. The reasons for the low adoption of AFS are discussed.

Key words: climate change; sustainability; forest recovery; biological controls; silvopasture.

RESUMEN

Los sistemas agroforestales (SAF) son policultivos que incluyen por lo menos una especie arbórea. Los SAF ofrecen diversos servicios ecosistémicos, aumento de las ganancias y seguridad alimentaria. Los servicios ecosistémicos de los SAF incluyen conservación de la biodiversidad, captura de carbono, reducción de las enfermedades de los cultivos, aumento de los controles biológicos, fijación biológica del nitrógeno y reciclaje de nutrientes. En este artículo se presenta una revisión de los servicios ecosistémicos asociados a los SAF en el sur del Brasil. Además se estima la potencialidad para la captura de carbono de los SAF. Los SAF predominantes son los asociados a la yerba mate (*Ilex paraguariensis* A. St. Hil.), sistemas silvopastoriles, cítricos, plátano y sistemas con la palmera açai-juçara (*Euterpe edulis* Mart.). Teniendo en cuenta la conversión de sistemas convencionales para SAF, el silvopastoreo presenta la mayor potencialidad de captura de carbono en función de la gran área de creación de ganado. La conversión de los cítricos y banana también presentan gran potencialidad de captura de carbono, además de la reducción de las enfermedades fúngicas y bacterianas. El sur de Brasil presenta más de 15 millones de hectáreas que pueden ser convertidas en SAF teniendo como modelo las experiencias ya bien establecidas. Además, los SAF están siendo adoptados como estrategia de restauración forestal. No se observan relaciones costo-beneficio negativas relacionadas con la adopción de silvopasturas o SAF con cítricos. Las razones para su baja adopción son discutidas.

Palabras clave: cambio climático; sustentabilidad; recuperación de bosques; controles biológicos; silvopastura.

¹ Submitted July 06, 2017 – Accepted November 09, 2017. This work is licensed under a [Creative Commons Attribution 4.0 International License](https://creativecommons.org/licenses/by/4.0/)

INTRODUCTION

Agroforestry Systems (AFS) are deliberated consortia of trees with crop plants and/or livestock, in determined space arrangements and sequences along the time, presenting varied interactions among their components (Baumer, 1991; Coelho, 2012). Besides economic and social advantages, the AFS provide ecological benefits such as erosion contention (Franco *et al.*, 2002; Traore *et al.* 2004), increment in the organic carbon in the soil and in the above-ground biomass (Albrech and Kandji, 2003; Mutuo *et al.*, 2005; Verchot *et al.*, 2007; Nair *et al.* 2009), biodiversity conservation (McNeely and Schroth, 2006; Harvey and Villalobos, 2007; Fajardo *et al.* 2009; Rivera *et al.*, 2013) and the promotion of spontaneous biological control associated with increased yields (Maas *et al.*, 2013). Moreover, agroforestry is agriculture practices which can generate higher ecologic sustainability than the conventional practices, whether organic or not. Agroforestry aggregates ecological functions such as soil erosion control, nitrogen leaching reduction and carbon uptake for which organic agriculture does not reach differentiation in relation to the conventional agriculture (Wilson and Lovell, 2016).

This paper is aimed at present the traditional and innovative AFS in Southern Brazil, as well as the reported or potential ecosystem services they provide. As the ecosystem services that could be provided by the AFS compound a wide suite of possibilities among cultural, environmental, social and economic perspectives (Fagerholm *et al.*, 2016), only the ecosystem services related to the biophysical aspects were addressed. Such circumscription included climate change and carbon sink, biological controls, reduction of agrochemicals and fertilizers and biological conservation. Economic aspects related to them are also commented. This delimitation is justified also by virtue of the available scientific background based on the region: as far as possible, the review is mainly based on data reports from Southern Brazil in itself and complemented with investigations from other Brazilian regions and even from other parts of the world whenever necessary.

The Southern region of Brazil includes the States of Paraná (PR), Santa Catarina (SC) and Rio Grande do Sul (RS). The humid subtropical climate (Cfa and Cfb in the Köppen-Geiger system) presents an annual average temperature and annual rainfall ranging between 14° to 22 °C and 1,250 to 2,000 mm, respectively (Leite, 1995). The pristine vegetation cover was predominantly forests of the Atlantic Forest Biome and a minor portion of grasslands or savannah-like ecosystems in the southernmost Rio Grande do Sul

State, which corresponds to the Pampa Biome (Figure 1).

In the Pampa Biome, the arboreal components are restrained to the gallery forests or as sparse components of woodlands or steppe-savannah complex, particularly at the extreme Southwest. Nevertheless, the highly anthropized grasslands in the Pampa Biome coexist with woody formations in a metaclimax dynamics since the current climate is favorable for both (Pillar and Vélez, 2010). Thus, the integration between livestock and silviculture exhibits high feasibility in this Biome (Saibro *et al.*, 2009).

With its favorable climate for the development of forests, the southern region of Brazil is highly auspicious to the development of agroforestry systems. However, due to historical and cultural factors, these systems remained limited to small areas until now. Meanwhile, in a context of growing environmental and social concerns about modern forms of agricultural managing, AFS have been seen as safer and more sustainable mode of production with diversified environmental benefits or services (Garrity, 2004, Coelho, 2012, Nair and Garrity, 2012).

The predominant agroforestry systems in Southern Brazil are yerba mate (*Ilex paraguariensis* A. St. Hil.), plantations and extractivism, *Citrus* spp. orchards, banana cultivation, açai-juçara (*Euterpe edulis* Mart.), silvopastures, and in a minor extent coffee systems. Hereafter the key features, the occupied area, the biophysical ecosystem services (reported or potential) and economic aspects are presented and discussed.

Agroforestry systems with yerba mate (*Ilex paraguariensis*)

The yerba mate is a native tree species from Brazil, Argentine and Paraguay, whose cultivation dates back to early European settlement in South America in the seventeenth century, although its extractive use is pre-Columbian (Linhares, 1969, Lagier, 2008). Similar to cocoa and coffee, yerba mate plants are shade-tolerant (Coelho and Mariath, 1996, Coelho *et al.*, 2011) and its evolution occurred amid the forests of the southern portion of the Atlantic Forest Biome. Although yerba mate could be cultivated at full sunlight with high densities as a monocrop, there are systems in which yerba mate is kept partially shaded under remaining native trees, and in which its density can be gradually increased (Figure 2). The shading could change the chemical composition (Coelho *et al.*, 2007a) and taste of yerba mate products (Streit *et al.*, 2007; Pagliosa *et al.*, 2009), and there is a common sense that this change is positive. Thus, the industry tends to pay higher values for raw materials coming from shaded systems. The increased appreciation of shading has promoted

the introduction of shading trees in mate plantations, in many cases native species, or the management of spontaneous growth of woody species, for instance *Araucaria angustifolia* (Bert.) O. Kuntze (Figure 2). Among the introduced shading species are the leguminous tree species *Mimosa scabrella* Benth and *Ateleia glazioveana* Baill.

Ateleia glazioveana is a deciduous species (Figure 2) whose pruning material can be used for mulching, presenting better results than animal manure with the same N contents (Baggio and Soares, 2006). In wild areas, the association between *A. glazioveana* e *I. paraguariensis* is noteworthy, which could involve some kind of facilitation (Coelho *et al.*, 2011). In addition, *A. glazioveana* could be introduced by direct seeding, with low costs (Escaio *et al.*, 2012).

The biological nitrogen transference from the atmosphere to vegetation is a key process in the

secondary succession, being the main limiting nutrient in the early phases (Amazonas *et al.*, 2011). Apart from other benefits such as firewood production and windbreak effect, the use of leguminous trees can reduce the dependence of industrial fertilizers, which constitutes an ecosystem service in itself (Kremen and Miles, 2012). Traditional systems with intercropping and rotation between maize and *M. scabrella* in Paraná State showed 82 kg ha⁻¹ of nitrogen surplus after each cycle of six years (Somarrriba and Kass, 2001).

Yerba Mate AFS reached a 63 Mg.ha⁻¹ of carbon stock on the aboveground biomass (Bastos, 2013), which is equivalent to forest remnants in intermediary successional stages, at the same locations. The plant biodiversity also reached similar values to forest remnants, though the floristic composition differs (Bastos, 2013), possibly caused by a selective management carried out by landowners.

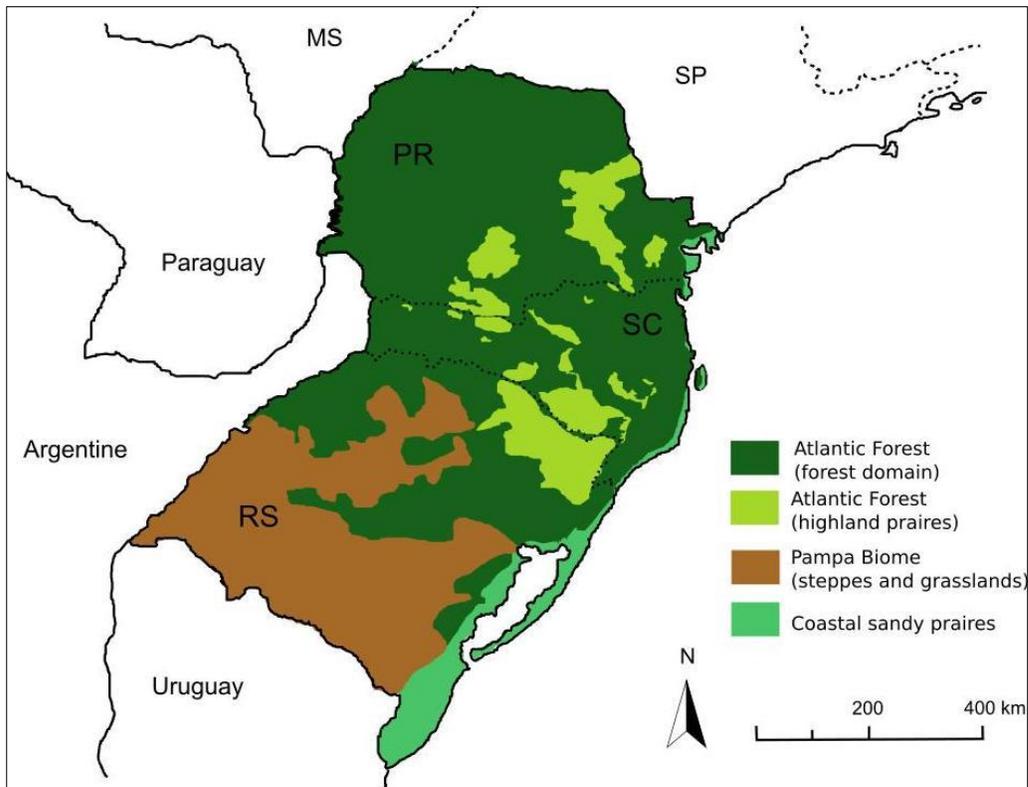


Figure 1. Pristine vegetation cover of Southern Brazil. The Atlantic Forest Biome included the Forest Domain and the Highland Prairies. The Pampa Domain includes the Steppes and Grasslands at the south and southwest, which extends to Uruguay and Argentina.



Figure 2. Agroforestry systems with yerba mate (*I. paraguariensis*). A. forest converted to high density yerba mate (in the foreground and amid the higher trees) cultivation with remaining wild trees in Machadinho, RS, 2007; B. agroforestry with yerba mate and *A. glazioviana* (timbó) in the superior strata in Chapecó, SC, 2013; C. agroforestry with yerba mate and *A. angustifolia* associated with pasture in Chapecó, SC, 2010; D. consortia with yerba mate, *M. scabrella* (bracatinga) and dairy cattle in Augusto Pestana, RS, 2009.

In Brazil, mate cultivation is virtually exclusive to the Southern States occupying 77,340 hectares (Table 1; IBGE, 2009, 2016). The total annual production of raw material is around 660 Gg and 1/3 is provided from agroforestry (Signor, 2013). Nevertheless, the total area maintained in agroforestry systems (managed forests + converted yerba mate plantations) is unknown. Assuming that 2/3 of the mate cultivation can be converted into agroforestry, and according to available reports (Bastos, 2013), near 1.0 Tg C could be incorporated in the aboveground biomass. The data regarding the potential for incorporating carbon in belowground biomass are not available (this point will be discussed forward).

On the other hand, the expansion of yerba mate cultivation through conversion of undisturbed forests

or those at an advanced successional stage can represent a liquid transference of carbon to the atmosphere. Mature undisturbed forests in Southern Brazil can contain more than 150 Mg ha⁻¹ C on the aboveground biomass (Mognon *et al.*, 2013). Thus, a simplification of the ecosystem to levels that allow a satisfactory productivity of yerba mate could represent losses of approximately 80-100 Mg C ha⁻¹.

Despite the evidence that AFS generally reduce pest incidence and pesticide needs (Steffan-Dewenter *et al.*, 2007; Maas *et al.*, 2013), the few available data from yerba mate systems indicate a similar level of herbivory in both monocropping and agroforestry systems (Avila *et al.*, 2016). Further analyses are needed to validate these findings.

Table 1. Area of the three Brazilian Southern States and the total area occupied by the agricultural activities which present the more expressive agroforestry or silvopasture experiences, and also with higher potential of conversion to agroforestry. RS = Rio Grande do Sul State, SC = Santa Catarina State, PR = Paraná State, BR(%) = percent of the area in relation to the total area of the country. According to IBGE (2009, 2016).

	RS	SC	PR	Southern region (total)	BR(%)	Brazil
Area (km ²)	281,748	95,346	199,315	576,409	6.77	8,515,767
Pastures	9,206,664	1,701,519	4,702,546	15,610,729	9.75	160,042,064
Citrus	39,798	28,107	37,459	105,364	12.34	854,010
Banana	12,226	29,534	11,000	52,760	10.10	522,300
Açaí/Palm heart	6	3,301	1,108	4,415	8.97	49,242
Yerba Mate	35,240	13,467	28,629	77,336	99.62	77,630
Coffee	--	--	38,333	38,333	0.19	2,002,151
Total	9,293,934	1,775,928	4,780,742	15,850,604	9.81	162,093,457

Table 2. Major ecosystem services contributions from the agricultural activities which present the more expressive agroforestry or silvopasture experiences and higher potential of conversion to agroforestry in Southern Brazil.

	Carbon uptake and stocking	Pesticide reduction	Biodiversity conservation	Decrease of industrial fertilizers dependence and biological fixation of N	Erosion control and watershed protection
Pastures	High	?	Low to High ²	Mid to High ²	High
Citrus	High	High	High	High	High
Banana	High	High	High	High	High
Açaí/Palm heart	Managing dependent ¹	?	High ⁴	?	Managing dependent ¹
Mate	Managing dependent ¹	Low? ³	High	High	Managing dependent ¹
Coffee	Mid to High	Mid to High	High	High	?

¹The conversion of monocropping into agroforestry systems can increase the ecosystem services, while the conversion of forests to agroforestry with yerba mate or açaí may produce the opposite effect.

²Dependent of the species used; for example, there are frequent silviculture of *Eucalyptus* (and other exotic species) with pastures that do not present high native diversity or leguminous trees with biological fixation of nitrogen.

³Based in the only reference available (Avila *et al.*, 2016).

⁴Considering a good conservation of canopy diversity.

Citrus and other fruit cultivation

Different *Citrus* species can be cultivated under the canopy of shading trees (Figure 3). The tree species of the upper stratum can provide benefits for the cultivation in several ways: biological fixation of nitrogen, nutrients cycling, protection against weather stresses. At shaded citrus orchards, the incidence and severity of typical diseases such as the bacteria *Xanthomonas axonopodis* (citrus canker) and the fungi *Guignardia citricarpa* (citrus black spot) have been significantly decreased (Gonzatto, 2009; Gonzatto *et al.*, 2010). On the other hand, the yields are not affected by moderate shading (Syvertsen *et al.*, 2003; Cohen *et al.*, 2005) or even increase (Gonzatto, 2009). Species like *Citrus sinensis* Osbeck and *C. limon* (L.) Osbeck show photosynthetic saturation at 30-40 % of the full sunlight, and temperatures over 20-30 °C

(depending on humidity) can inhibit the photosynthesis (Kriedemann, 1968; Wheaton *et al.*, 1978; Jifon and Syvertsen, 2003). The inhibition of the citrus canker could be related to the windbreak effect of the associated trees (Tamang *et al.*, 2010), which reduces the damage of leaves.

However, agroforestry can promote higher infestation of citrus blackfly *Aleurocanthus woglumi* Ashby, 1915 (Sternorrhyncha: Aleyrodidae) compared with conventional cultivation (Silva *et al.* 2011a), although the difference on damages was not evaluated. The citrus blackfly was more commonly associated with the warmer regions of Brazil and the first observation was in Pará State, Amazonas region; the occurrence in Southern region is rare until this moment (Molina *et al.*, 2014).

Shading increases the longevity of the *Citrus*, which maintains high productivity for a longer time span. According to landowner's information, citrus plants in shaded orchards from the Rio Grande do Sul State provide high yields even after 25 years.

Despite the strong evidence of economic and ecological benefits of agroforestry over conventional orchards, this form of cultivation is still an exception. Brazil has more than 850.000 ha of citrus plantations (Table 1). In the absence of reliable inventories, one could estimate that *Citrus* agroforestry performs less than 2-3 % of the total area. The inexistence of tradition, cultural resistance and lack of knowledge among the rural extension agents are explanatory factors for this situation.

Conversion of *Citrus* orchards to agroforestry can incorporate carbon in expressive amounts. Available estimation in the region indicates values around 25 Mg C ha⁻¹ in the aboveground biomass (20 % is from *Citrus* plants), which corresponding to 50 % of the aboveground biomass at forest remnants in pairwise comparisons in the same region (Bastos, 2013). The government census (IBGE, 2009, 2016) estimates 105,000 ha of citrus cultivation area at the Southern region, thus the potential to incorporate C on the aboveground biomass is around 2.1 Tg, e.g. an addition of 20 Mg C ha⁻¹. On the other hand, the data regarding the potential for incorporating carbon in belowground biomass are not available.



Figure 3. Agroforestry systems with fruit cultivation. A. agroforestry with *Citrus* with a row of shading native trees arranged by the cutting management in Aratiba, RS, 2015; B. agroforestry with açai-juçara (*Euterpe edulis* Mart.) in the understory and an individual of *Plinia peruviana* (Poir.) Govaert in the foreground on the left in Barra do Turvo, São Paulo State, 2009; C. agroforestry with *Citrus* with *Parapiptadenia rigida* (Benth.) Brenan in the canopy (a leguminous tree with biological fixation of nitrogen) in Tupandi, RS, 2009; D. banana agroforestry with *Machaerium stipitatum* Vogel (another leguminous tree with BFN) in Dom Pedro de Alcântara, RS, 2009.

The banana cultivation also presents benefits when moderately shaded, eventually showing higher productivity with mild shading. The beneficial effect of shading increases inversely with the latitude (Norgrove, 1998). The yellow Sigatoka (*Mycosphaerella musicola*), one of the main diseases that affect this crop, is reduced significantly with the presence of the shading trees in agroforestry (Norgrove and Hauser, 2013), a fact also observed in Southern Brazil, according to reports from producers. The reduction of the yellow Sigatoka and also of the black Sigatoka (*Mycosphaerella fijiensis*) observed in Brazil can occur due to different factors, such as windbreak effect and the consequent reduction of foliar damage, better nutrient cycling, and the reduction of leaf surface temperature, which impair the fungus development; however, excessive shading can increase Sigatoka incidence, possibly by increased humidity (Favreto *et al.*, 2007).

The açai-juçara (*Euterpe edulis* Mart.) is a crop with growing value in Southern Brazil, with annual expansion rates of over 7 % (IBGE, 2009). It is native of the Atlantic Forest Biome and presents a high shade tolerance, growing spontaneously in the dense forests along the coast and in the middle Paraná River basin, among Brazil, Paraguay and Argentine border. Traditionally, extrativism has focused on the meristematic apex, the 'heart of palm' or 'palmito'. However, extraction of the stem apex implicates the loss of the plant, which has conducted the species to the edge of extinction. Nowadays, their fruits reach higher commercial value, preserving the stem and the plant. As a shade-tolerant species, *E. edulis* is very well adapted to the agroforestry regime (Figure 3). On the other hand, as an endangered species, a great challenge to producers is to comply with the restrictions in the Brazilian environmental laws for its cultivation and commerce (Chaimsohn and Chiquetto, 2013).

Supposedly, agroforestry with *E. edulis* can contribute to the biological conservation, if high canopy diversity is maintained. However, the dynamics of biodiversity associated with açai-juçara and consequently the very contribution to ecosystems services from this crop is poorly investigated.

Silvopastures

The integration between trees and pastures or simply silvopastures (Figures 2, 4) is possibly the agroforestry practice most prevalent in Brazil. Beef or dairy cattle achieve higher animal comfort and productivity (Yamamoto *et al.*, 2007; Paciullo *et al.*, 2011), which is also valid for the sheep (Magalhães *et al.*, 2001). Regarding the ecosystem services, silvopastures can contribute to carbon uptake, nutrient cycling, erosion control, biodiversity conservation and reduce the dependence of external inputs (Murgueitio *et al.*,

2011). In addition, many tree species are good quality foragers with high levels of protein and minerals (Mpairwe *et al.*, 1998; Datt *et al.*, 2008; Santos *et al.*, 2010; Perez *et al.*, 2013).

Brazil has 160 million hectares of pastures (IBGE, 2009). By assuming a potential of 4.6 Mg ha⁻¹ y⁻¹ of carbon uptake in aboveground biomass (Kim *et al.*, 2016), it means a total potential uptake or 0.74 Pg C year⁻¹, which corresponds to 61.7 % of Brazilian annual carbon emissions. The southern region of Brazil takes 9.8 % of the total Brazilian pastures (Table 1). The rate of carbon uptake in new silvopastures and other agroforestry systems can remain at this level for at least 25 years (Kim *et al.*, 2016), which can constitute a propitious contribution while a transition in energy matrix sources takes place in order to reduce carbon emission.

Carbon uptake can be more than doubled by taking into account the belowground carbon. Agroforestry systems can accumulate 24 Mg C ha⁻¹ in the soil over the conventional croplands considering a 0.4 m depth (Maia *et al.*, 2007). However, the soil carbon contents in silvopastures could be even higher below 0.75 m (Haile *et al.*, 2008), and the carbon uptake from conversion to agroforestry systems can surpass 100 Mg C ha⁻¹ including levels below 1.0 m (Makumba *et al.*, 2007). The deeper development of the tree roots in relation to annual herbaceous crops can explain this increment (Lorenz and Lal, 2014), the tree roots connecting atmosphere to subsoil as a carbon transfer path. Moreover, root-derived carbon is retained in soils much more efficiently than are above-ground inputs of leaves and needles (Schmidt *et al.* 2011). In such wise, summing the carbon uptake rates of above and below ground biomass and of soil, a value of 14.0 ± 4.1 Mg ha⁻¹ y⁻¹ can be registered for silvopastures (Kim *et al.* 2016). Notwithstanding, Udawatta and Jose (2012) estimated an uptake of 6.1 Mg ha⁻¹ y⁻¹ for silvopastures in North America. Biomass stocks and increment is highly variable, which points out the importance of regional inventories and reliable methodologies (Agevi *et al.*, 2017).

Other agroforestry experiences

The aforementioned cases are the more expressive in terms of area and economic influence. However, it is important to address other agroforestry cases at Southern Brazil, such as agroforestry with coffee plantations. Despite the reduction of planted area with coffee in Paraná State (from 112.000 ha in 2006 to 38.000 ha in 2014) (IBGE, 2009, 2016), virtually the only State with commercial plantations in the Southern Brazil, the agroforestry with this crop is feasible in the region, considering yields and economic return when compared with monocropping (Baggio *et al.*, 1997). The contribution of shading to flavor is a matter of

controversy (Rapidel *et al.*, 2015) but some attempts to aggregate value to the coffee from organic agroforestry has been recently carried out in Paraná State (Bronzeri and Bulgacov, 2014). The reduction of pests is also controversial and it can be affected by local factors (Rapidel *et al.*, 2015). However, the use of leguminous shading trees such as *M. scabrella* (Caramori *et al.*, 1996) could offer a significant contribution to the BFN and for reduction of use of industrial fertilizers application.

Horticulture including annual or biannual crops also has been tested in the agroforestry mode here and there. Among these relatively isolated experiences, pineapple is an outstanding case (Figure 4), considering that a mild shading can benefit this crop and pineapple fruits are sensitive to excessive light and temperature in their maturation stage (Liu and Liu, 2012). At agroforestry, Brazilian producers report an expanded period of pineapple fruit maturation than in the full sunlight cultivation, which takes an advantage due to higher commercial values achieved off-season.

Agroforestry as an Ecological Restoration Strategy

Recently, the agroforestry has been utilized in Brazil as a strategy of ecosystem restoration (Vieira *et al.*, 2009), overcoming the paradigm of an inherent conflict between agriculture and ecological restoration and conservation (Figure 4D and 4E). Moreover, the growth of the tree saplings could be higher in the agroforestry than in seedling plantations followed by abandon (Coelho, 2010). The explanation for this could be due to the inhibition of the trees by the herbaceous heliophilous plants, mainly Poaceae (Souza and Batista, 2004; Yu, 2004; Pompéia, 2005; Ziller *et al.*, 2010). In AFS these competitive plants are controlled by the intercropping cultivation, at least during the first years. In addition, fertilizer residues from intercropping cultivation can be captured by tree roots. Moreover, the restoration costs can be partially amortized with the crop yields, similarly to the Taungya system (Rodrigues *et al.*, 2008), in which an intercropping of short cycle crops with timber species can promote a faster economic return and higher cash flow in the first years.

Ecosystem services of the Agroforestry in Southern Brazil

Biological nitrogen fixation

Improvement of nitrogen use efficiency in agriculture can be considered a relevant ecosystem service, face to the growth of human population and food supply demand (Spiertz, 2010). Agroforestry practices can improve the N use efficiency in different ways. First, the trees through their deep root systems can capture the N that would otherwise be lost to the groundwater and the atmosphere (Kremen and Miles, 2012). Secondly, microorganisms associated with plants can transfer N from the atmosphere to the trophic chain through biological fixation. As perennial elements, the trees in the agroforestry systems could represent a low-cost biological N fixation, capturing on average 250 (56-675) kg N ha⁻¹ year⁻¹ (Nygren *et al.*, 2012).

Somarriba and Kass (2001) studied a six-year cycle of a traditional rotation agroforestry with *Mimosa scabrella* at Paraná, Southern Brazil, verifying 356 kg ha⁻¹ of N added to the soil and 13.7 kg ha⁻¹ year⁻¹ of N surplus considering the aboveground biomass (accumulation minus exportation). Certain symbiotic N₂-fixing bacteria strains associated with *M. scabrella* can provide all N required by the plant (Primieri *et al.*, 2016). Field evaluation indicates that 90 % of accumulated N in *M. scabrella* is derived from biological N fixation (Coelho *et al.*, 2007b).

Several other Fabaceae tree species from Southern Brazil are promising N fixers, standing out those from the genera *Enterolobium*, *Albizia*, *Ateleia*, *Erythrina*, *Machaerium*, *Inga*, *Mimosa*, *Parapiptadenia*, and *Vachellia* (= *Acacia* p. p.). For instance, *Parapiptadenia rigida* (Benth) Brenan and *A. glazioveana* have been associated with citrus and yerba mate cultivation. The second species present roughly 3.1 % of N content and is well adapted to acid soils with high Al levels (Baggio *et al.*, 2002). *P. rigida* presents a lower N content (2.1 % according to Zanella and Coelho, 2014) although is commonly used to shading citrus (Figure 3C) and is very well adapted to acid and rocky soils.

N fixing effectiveness and efficacy of native leguminous trees from southern Brazil need further investigation in order to promote the economic use and conservation of native biodiversity.



Figure 4. Silvopasture and other agroforestry systems at Southern Brazil. A. Silvopasture with remnant native trees in Guatambu, SC, 2015; B. Agroforestry with pineapple, orange and banana with native shading trees in Nova Laranjeiras, PR; C. beans intercropping with *Trema micrantha* (L.) Blüme and *Mimosa scabrella* Bentham in Catuípe, RS, 2007; D-E. agroforestry for forest restoration with cassava, maize and pumpkins associated with native trees such as *Handroanthus albus* (Cham.) Mattos and *Peltophorum dubium* (Spreng.) Taub. on left, *Cedrela fissilis* Vell. and *Heliocarpus americanus* L. on right, in Doutor Maurício Cardoso, RS, 2003.

Carbon sequestration with agroforestry

Carbon sequestration via AFS presents low costs when compared with other strategies (Verchot *et al.* 2005) and can incorporate 2.0-5.8 Mg C ha⁻¹ year⁻¹ (Concha *et al.*, 2007). If compared to wild rainforests or afforestation (Allen *et al.*, 2009), AFS can maintain between 60-80 % of the methane absorption capability, which could be explained by the reduction of the bulk soil density and the increase of porosity and O₂ availability in soils (Mutuo *et al.*, 2005).

The soil changes promoted by the agroforestry include an increment in the amount of total organic matter and recalcitrant organic matter ((Hawke and O'Connor, 1993; Muñoz *et al.*, 2007; Haile *et al.*, 2008). Concomitantly, AFS reduce pH due to cation transference from the soil to the aboveground biomass and/or increased levels of acidic organic matter

(Hawke and O'Connor, 1993; Sharma *et al.*, 2009). AFS increase the nutrient cycling due to growing litter production, which is related to tree aging and basal area (Bhojvaid and Timmer, 1998; Kumar, 2008). On the other hand, it should be stressed that litter production in agroforestry could change by a magnitude of 20 depending on the tree species, and the pioneer and early secondary species present the higher values (Benvenuti-Ferreira *et al.*, 2009). Nutrient cycling is also enhanced as a function of a higher microbiological activity under AFS (Vallejo *et al.*, 2010).

Aforementioned changes in the soil observed in AFS resembles those modifications observed along the aging of afforested sites (Berthrong *et al.*, 2009; Wen-Jie *et al.*, 2011) and of secondary spontaneous succession in subtropical and tropical forest ecosystems (Feldpausch *et al.*, 2004; Coelho *et al.*,

2011; Schwiderke *et al.*, 2012) since the growth of trees is the key driver of the changes in the soil properties in these ecosystems. The presence of shading trees also contributes expressively to reducing erosive processes (Rodríguez and García, 2009; De Aguiar *et al.*, 2010), which constitutes one of the main ecosystem services of agroforestry, for example the protection of water sources and soil fertility. Moreover, AFS tends to present great soil porosity and water retention (Silva *et al.*, 2011b), promoting a more sustainable use of water resources.

CH₄ and N₂O greenhouse gases emissions

Agroforestry potential for reduce the emission of methane from ruminating animals due to the reduction of heat stress and the increase in pasture quality has been hypothesized. Preliminary results obtained at Southern Brazil (Pontes *et al.*, 2014) indicate a reduction of near 40 % in methane emission, although the difference between treatments was not statistically significant due to the small sample. On the other hand, in experiments in controlled stable conditions an inverse correlation between methane emission and temperature in a 5-20 °C interval was observed (Ngwabie *et al.*, 2011). Mechanistic evolving methane emission is highly complex (Allard *et al.*, 2007; Knapp *et al.*, 2014) and further field investigations are needed to validate whether the silvopastures could present such additional effect in the carbon cycle. Notwithstanding, silvopastoral systems can reduce methane emissions as a result of changes in the physicochemical properties of the soil, which are promoted by the presence of trees (Allard *et al.*, 2007; Knapp *et al.*, 2014). The increase of the soil porosity in agroforestry (and also in forests and riparian buffers) is a key factor to increment the CH₄ oxidization, producing a net uptake from the atmosphere (Rowlings *et al.*, 2012; Kim *et al.*, 2016).

However, these effects are not yet sufficiently studied and a definitive explanation is not yet available (Allen *et al.*, 2009). Further investigations are urgently needed considering that the direct emissions from the cattle reach 64 % of the emissions from the agriculture in Brazil (Brasil, 2016).

The influence of agroforestry adoption on the emission of N₂O present conflicting results and no significant difference to other agricultural land uses was observed (Kim *et al.*, 2016). On the other hand, N₂O emissions from agroforestry reported by these authors (1.3 to 10.1 kg N₂O ha⁻¹ year⁻¹) are in the range of tropical and subtropical forest emissions (5.1-74.5 kg N₂O ha⁻¹ year⁻¹; Rowlings *et al.*, 2012), which indicates that agroforestry *per se* does not enhance N₂O release for atmosphere. The increment in the N₂O emissions in agroforestry could be associated with the biological N fixation by leguminous trees (Kim *et al.*, 2016).

However, comparisons between biological N fixation and synthetic N fertilizers as a source of N in agriculture indicate that the synthetic sources present higher N₂O emissions (Bayer *et al.*, 2015). Again, agroforestry in itself could not be the driver of N₂O emission elevation. N management (leguminous trees, residues incorporation, green and animal manure) would be the focus for strategies of N₂O reduction.

Agroforestry and biological control

When compared to monocultures, agroforestry tends to reduce weeds, disease and herbivory. However, results are highly context-dependent. Factors such as pest and tree species identity and crop type may play a major role (Schroth *et al.*, 2000; Pumariño *et al.*, 2015). Interplanting of *Citrus* with *Psidium guajava* at Vietnam reduced the incidence of the bacteria *Candidatus liberibacter* (greening disease) only for one year, after what the effect was null (Ichinose *et al.*, 2012). However, interplanting with other fruit trees of the same stature should not be considered equivalent to the agroforestry orchards observed in Brazil (Figure 3A and 3C) where the shading trees perform a dossel over the citrus plants. Moreover, since the agroforestry citrus plantings in Southern Brazil are not isolated from other contaminated orchards, it could be hypothesized that a kind of increased resistance to the fungi and bacterial diseases is established. Similar beneficial effects were observed in cacao and coffee, for what microclimate modifications due to shade can control pathogenic fungi and bacteria (Avelino *et al.*, 2011). In spite of the relative high number of studies on the relationship between agroforestry management and disease and insect pests (Philpott and Ambrecht, 2006; Tschardtke *et al.*, 2011), investigations on citrus and yerba mate agroforestry are surprisingly scarce.

On the other hand, the biological control effectiveness on agroforestry could be related to factors of landscape scale. For example, distance from remnant forest patches can interfere with the pest and enemies populations (Tschardtke *et al.* 2008, De la Mora *et al.* 2015). Some taxa may be more sensitive to landscape effects. For instance, Lepidoptera increased in abundance on sites located at higher distances from the primary forest in Cacao agroforestry (Maas *et al.*, 2013) but different Lepidoptera species could present opposite responses to landscape variables in coffee agroforestry (Muriel *et al.*, 2014). Notwithstanding, landscape effect biological controls and pest and disease incidence on agroforestry is poorly studied around the world, and studies from Brazil are lacking.

Economic trade-offs of agroforestry adoption in Southern Brazil

Despite the environmental contribution, the economic balance is a key factor in the adoption of AFS. Broadly,

the economic advantages come from a higher productivity of the set of cultures, with an equivalent area ratio higher than 1.0, indicating complementarities in the use of resources by the different cultures and a low competition level (Van der Werf *et al.*, 2007; Martin and Van Noordwijk, 2009). In addition, economic advantages can come from indirect advantages obtained with costs reduction or quality increment, for instance through biological fixation of nitrogen, protection against climatic extremes or reduction of pests and diseases (Baggio *et al.*, 1997; Nygren *et al.*, 2012; Maas *et al.*, 2013), even when the main culture experiences a yield decrease.

In the yerba mate case, the yields do not are reduced in moderate shading (Coelho *et al.*, 2007a). However, the optimal shading level is little known and should vary between regions and growth conditions. Nevertheless, it is possible to introduce shading trees with high timber value, increasing the economic income without reducing yerba mate yields (Baggio *et al.*, 2011). The high trading value of the raw material from shaded cultivation can also aggregate value.

Regarding citrus, the moderate shading does not reduce and could even increase the productivity (Gonzatto, 2009). However, the greatest economic benefit of shaded orchards is the reduction of fungal and bacterial diseases. This also applies to the cultivation of bananas.

For silvopastoral systems, the situation is no different. The economic evaluations indicate that the silvopastures present higher economic incomes when compared to the forestry or conventional cattle farming (Paciullo *et al.*, 2011; De Souza *et al.*, 2015).

As an overall conclusion, there is no evidence of economic conflict in the adoption of agroforestry systems in Southern Brazil. As with other innovations, agroforestry adoption and permanence is influenced by several social, economic, and biophysical factors (Mercer, 2004). Land tenure, age, education level, self-efficacy, attitudes, social regulation, availability of credit and markets, labor resources, public policies (or lack thereof), among others, are the variables significantly associated with agroforestry adoption (Pattanayak *et al.*, 2003; McGinty *et al.*, 2008; Miccolis *et al.*, 2011; Meijer *et al.*, 2016). However, discrepancies among theoretical framework, methodologies, and selection of variables have led to a scientific puzzle (Mercer, 2004). A theoretical synthesis and even a rank of the relative importance of the different factors remain unavailable. In Southern Brazil, the public policies and government initiatives towards agroforestry development are still scarce, and usually, the few official programs are restricted to consortia of *Eucalyptus* (or other exotic species) and cattle. Almost all the successful cases related here are

isolated developments derived from landowners' experience or projects of NGOs and small cooperatives. Nevertheless, they would be models for future expansions and research. For example, NGOs were able to establish innovative approaches such as participatory design and partnerships, fostering the technical improvement of agroforestry systems in the Northern Atlantic Forest (Cardoso *et al.*, 2001; Souza *et al.*, 2012). In the Rio Grande do Sul State, cooperatives and NGOs have promoted a greater commercial value of native fruit species in agroforestry systems (Tonin *et al.*, 2017).

Beyond the scarcity of public initiatives, restriction laws to the economic use of native species in Brazil (Coelho, 2012; Chaimsohn and Chiquetto, 2013), following the example of similar legal barriers in other countries (Detlefsen and Somarriba, 2015; Nath *et al.*, 2016), constitute an additional factor which can inhibit agroforestry expansion, at least high diverse agroforestry with native biodiversity. Other factors remain to be clarified considering the scarcity of studies on the adoption of agroforestry in Brazil.

CONCLUSIONS

Agroforestry systems can offer many ecosystem services, for example carbon uptake and global warming mitigation, biodiversity conservation, biological controls and reduction of pesticides application, erosion control, biological fixation of nitrogen and nutrient cycling, reducing the dependence of industrial fertilizers. Most of its services are related directly to soil changes: increase of carbon, porosity, and flux of nutrients. In the Southern region of Brazil several innovative initiatives on agroforestry, mostly designed by the own farmers, have demonstrated feasibility not only in terms of increasing yields or economic return, but also by providing ecosystem services. The main productive categories which present well-established experiences and also the great potential for the successful conversion from conventional cultivation to agroforestry systems in Southern Brazil are silvopastures, *Citrus* orchards, and banana plantations, mainly due to the great extent of these activities in Brazil. In both activities, ecological and economic advantages are convergent and recognized also at academic level. Other agroforestry systems such as yerba mate (*I. paraguayensis*) and açai-juçara (*E. edulis*) could also offer ecosystem services, despite the need of further investigations to elucidate some controversial questions, for instance the contribution to biological controls and the adequate level of shading. In addition, the impact of the conversion of remnant forests to agroforestry also deserves further attention. A great challenge to the region is to qualify the extension initiatives to improve the adoption of the agroforestry systems, at least for the crops to which there are better scientific bases.

Acknowledgements

To Paulo Roger Lopes Alves (Universidade Federal da Fronteira Sul, Brazil) who reviewed a preliminary version of this paper. To Fernando da Silva for the English review. To Santo Gabriel Vaccaro (Universidade Federal da Fronteira Sul, Brazil) for the resumen review. To the farmers who have welcomed and taught us about agroforestry. To CNPq for the research grant (477973/2012-4).

REFERENCES

- Agevi, H., Onwonga, R., Kuyah, S., and Tsingalia, M. 2017. Carbon stocks and stock changes in agroforestry practices: a review. *Tropical and Subtropical Agroecosystems*, 20: 101-109.
- Albrecht, A. and Kandji, S. T. 2003. Carbon sequestration in tropical agroforestry systems. *Agriculture, Ecosystems & Environment*, 99: 15-27.
- Allard, V., Soussana, J. F., Falcimagne, R., Berbigier, P., Bonnefond, J. M., Ceschia, E., D'Hour, P., Hénault, C., Laville, P., Martin, C. and Pinares-Patino, C. 2007. The role of grazing management for the net biome productivity and greenhouse gas budget (CO₂, N₂O and CH₄) of semi-natural grassland. *Agriculture, Ecosystems & Environment*, 121: 47-58.
- Allen, D. E., Mendham, D. S., Cowie, A., Wang, W., Dalal, R. C. and Raison, R. J. 2009. Nitrous oxide and methane emissions from soil are reduced following afforestation of pasture lands in three contrasting climatic zones. *Soil Research*, 47: 443-458.
- Amazonas, N. T., Martinelli, L. A., Piccolo, M. C. and Rodrigues, R. R. 2011. Nitrogen dynamics during ecosystem development in tropical forest restoration. *Forest Ecology and Management*, 262: 1551-1557.
- Avelino, J., Ten Hoopen, G. M., and DeClerck, F. 2011. Ecological mechanisms for pest and disease control in coffee and cacao agroecosystems of the neotropics. In: Rapidel, B., Jean-François, L. C., and Beer, J. (eds.) *Ecosystem Services from Agriculture and Agroforestry Measurement and Payment*. Earthscan, London, p. 91-117.
- Avila Jr, R. S., Dalazen, D. F., Lorentz, L. H., Poletto, I. and Stefenon, V. M. 2016. Effects of different cultivation systems in leaf traits and herbivory damage in *Ilex paraguariensis* (Aquifoliaceae). *Brazilian Journal of Botany*, 39: 219-223.
- Baggio, A. J., Caramori, P. H., Androcioli-Filho, A. and Montoya, L. 1997. Productivity of southern Brazilian coffee plantations shaded by different stockings of *Grevillea robusta*. *Agroforestry Systems*, 37: 111-120.
- Baggio A. J., Felizari, S. R., Ruffato, A., and Soares, A. O. 2011. Produção do componente arbóreo no sistema agroflorestal da erva-mate (*Ilex paraguariensis*) em Machadinho, RS. 5th Congresso Sudamericano de la Yerba Mate, Posadas, Argentine, May 5-6, 2011. Proceedings, Posadas: INYM/UNaM/INTA, pp. 107-118.
- Baggio, A. J., Montoya, L. J. V., and Masaguer, A. 2002. Potencialidad del timbó (*Atelesia glazioveana*) y el maricá (*Mimosa bimucronata*) para producción perenne de abonos verdes en zonas de clima subtropical. I - Persistencia y productividad. *Revista Investigación Agraria*, v. 17, n. 1, p. 101-112.
- Baggio, A. J., and Soares, A. O. 2006. Efeito da aplicação de *mulching* de timbó no desenvolvimento inicial da erva-mate. Proceedings of 1st Congresso Internacional Erva-mate, 4nd Congresso sudamericano de la Yerba Mate, 4nd Reunión Técnica de la Yerba Mate, Posadas, Argentine, November 5-8, 2006. Posadas: INYM/INTA/UNaM/EPAGRI, p. 257-262.
- Bastos, J. R. 2013. Estrutura e estoques de carbono em sistemas agroflorestais e remanescentes florestais no alto rio Uruguai, Sul do Brasil. Master Dissertation, URI – Universidade Regional Integrada, Erechim, Brazil.
- Baumer, M. 1991. Animal production, agroforestry and similar techniques. *Agroforestry Abstracts*, 4: 179-198.
- Bayer, C., Gomes, J., Zanatta, J. A., Vieira, F. C. B., de Cássia Piccolo, M., Dieckow, J., and Six, J. 2015. Soil nitrous oxide emissions as affected by long-term tillage, cropping systems and nitrogen fertilization in Southern Brazil. *Soil and Tillage Research*, 146: 213-222.
- Benvenuti-Ferreira, G., Coelho, G. C., Schirmer J., and Lucchese, O. A. 2009. Dendrometry and litterfall of neotropical pioneer and early secondary tree species. *Biota Neotropica*, 9: 65-71.
- Berthrong, S. T., Jobbágy, E. G., and Jackson, R. B. 2009. A global meta-analysis of soil exchangeable cations, pH, carbon, and nitrogen with afforestation. *Ecological Applications*, 19: 2228-2241.

- Bhojvaid, P. P., and Timmer, V. R. 1998. Soil dynamics in an age sequence of *Prosopis juliflora* planted for sodic soil restoration in India. *Forest Ecology & Management*, 106: 181-193.
- Brasil. 2016. Estimativas anuais de emissões de gases de efeito estufa no Brasil. 3rd Ed., Ministério da Ciência, Tecnologia e Inovação – MCTI, Brasília, Brazil. Available in < http://sirene.mcti.gov.br/documents/1686653/1706227/LIVRO_MCTIC_EstimativaDeGases_Publica%C3%A7%C3%A3o_210x297mm_FINAL_WEB.pdf/61e78a4d-5ebe-49cd-bd16-4ebca30ad6cd > accessed in June, 05, 2017.
- Bronzeri, M. S., and Bulgacov, S. 2014. Estratégias na cadeia produtiva do café no norte pioneiro do Paraná: competição, colaboração e conteúdo estratégico. *Organizações Rurais & Agroindustriais*, 16: 77-91.
- Caramori, P. H., Androciolli-Filho A., and Leal, A. C. 1996. Coffee shade with *Mimosa scabrella* Benth. for frost protection in Southern Brazil. *Agroforestry Systems*, 33: 205-214.
- Cardoso, I. M., Guijt, I., Franco, F. S., Carvalho, A. F., and Neto, P. F. 2001. Continual learning for agroforestry system design: university, NGO and farmer partnership in Minas Gerais, Brazil. *Agricultural Systems*, 69: 235-257.
- Chaimsohn, F. P., and Chiquetto, N. C. 2013. Construção do marco legal para a produção de Açaí de juçara: contribuições da “oficina interestadual sobre legislação, comercialização e marketing para exploração de frutos da palmeira juçara”. *Revista Conexão UEPG*, 9: 244-253.
- Coelho, G. C. 2010. Restauração florestal em pequenas propriedades: desafios e oportunidades. In: Hüller, A. (ed.) *Gestão Ambiental nos Municípios: Instrumentos e Experiências na Administração Pública*. FURI, Santo Ângelo, Brazil, pp. 195-215.
- Coelho, G. C. 2012. *Sistemas Agroflorestais*. Rima Editora, São Carlos, Brazil.
- Coelho, G. C., and Mariath, J. E. A. 1996. Inflorescences morphology of *Ilex* L. (Aquifoliaceae) species from Rio Grande do Sul, Brazil. *Feddes Repertorium*, 107: 19-30.
- Coelho, G. C., Rachwal, M. F. G., Dedeczek, R. A., Curcio, G. R., Nietsche, K., and Schenkel, E. P. 2007a. Effect of light intensity on methylxanthine contents of *Ilex paraguariensis* A. St. Hil. *Biochemical Systematics and Ecology*, 35: 75-80.
- Coelho, G. C., Rigo, M. S., Libardoni, J. B., Oliveira, R., and Benvenuti-Ferreira, G. 2011. Understory structure in two successional stage of Semi-deciduous Seasonal Forest remnant of Southern Brazil. *Biota Neotropica*, 11: 63-74.
- Coelho, S. R. de F., Gonçalves, J. L. de M., Mello, S. L. de M., Moreira, R. M., Da Silva, E. V., and Laclau, J. P. 2007b. Crescimento, nutrição e fixação biológica de nitrogênio em plantios mistos de eucalipto e leguminosas arbóreas. *Pesquisa Agropecuária Brasileira*, 42: 759-768.
- Cohen, S., Raveh, E., Lia, Y., Grava, A., and Goldschmid, E. E. 2005. Physiological responses of leaves, tree growth and fruit yield of grapefruit trees under reflective shade screens. *Scientia Horticulturae*, 107: 25-35.
- Concha, J. Y., Alegre, J. C., and Pocomucha, V. 2007. Determinación de las reservas de carbono en la biomasa aérea de sistemas agroforestales de *Theobroma cacao* L. en el departamento de San Martín, Peru. *Ecological Applications*, 6: 75-82.
- Datt, C., Datta, M., and Singh, N. P. 2008. Assessment of fodder quality of leaves of multipurpose trees in subtropical humid climate of India. *Journal of Forestry Research*, 19: 209-214.
- De Aguiar, M. I., Maia, S. M. F., Xavier, F. A. S., de Sá Mendonça, E., Araújo Filho, J. A., and De Oliveira, T. S. 2010. Sediment, nutrient and water losses by water erosion under agroforestry systems in the semi-arid region in northeastern Brazil. *Agroforestry Systems*, 79: 277-289.
- Detlefsen, G., and Somarriba, E. 2015. Producción agroforestal de madera en fincas agropecuarias de centroamérica. In: Montagnini, F., Somarriba, E., Murgueitio, E., Fassola, H. and Eibl, B. (eds.) *Sistemas agroforestales: funciones productivas, socioeconómicas y ambientales*, CIPAV, Cali, CATIE, Turrialba, pp. 21-43.
- De la Mora, A., Pérez-Lachaud, G., Lachaud, J. P., and Philpott, S. M. 2015. Local and landscape drivers of ant parasitism in a coffee landscape. *Environmental Entomology*, 44: 939-950.
- De Souza, Á. N., De Oliveira, A. D., Scolforo, J. R. S., De Rezende, J. L. P., De Mello, J. M. 2015. Viabilidade econômica de um sistema agroflorestal. *Cerne*, 13: 96-106.
- Escaio, A. C., Corrêa, G., Lima, J. D. N., and Coelho, G. C. 2012. Emergency and growth of *Ateleia glazioviana* Baill. seedlings in direct sowing

- in an early secondary succession stage. *Brazilian Journal of Ecology*, 14: 35-41.
- Fagerholm, N., Torralba, M., Burgess, P. J., and Plieninger, T. 2016. A systematic map of ecosystem services assessments around European agroforestry. *Ecological Indicators*, 62: 47-65.
- Fajardo, D., Neira, R., and Chará, L. 2009. Influencia de sistemas silvopastoriles en la diversidad de aves en la cuenca del río La Vieja, Colombia. *Recursos Naturales y Ambiente*, 58: 9-16.
- Favreto, R., Model, N. S., and Tonietto, A. 2007. Sigatoka Negra, fatores de ambiente e sistemas agroflorestais em bananais do Rio Grande do Sul, Brasil. *Pesquisa Agropecuária Gaúcha*, 13: 95-104.
- Feldpausch, T. R., Rondon, M. A., Fernandes, E., Riha, S. J., and Wandelli, E. 2004. Carbon and nutrient accumulation in secondary forests regenerating on pastures in central Amazonia. *Ecological Applications*, 14: 164-176.
- Franco, F. S., Couto, L., Carvalho, A. F., Jucksch, I., Fernandes Filho, E. I., Silva, E., and Meira Neto, J. A. A. 2002. Quantificação de erosão em sistemas agroflorestais e convencionais na zona da mata de Minas Gerais. *Revista Árvore*, 26: 751-760.
- Garrity, D. P. 2004. Agroforestry and the achievement of the Millennium Development Goals. *Agroforestry Systems*, 61: 5-17.
- Gonzatto, M. P. 2009. Desenvolvimento e produção de Citrus em sistema agroflorestal. Master Dissertation, UFRGS, Porto Alegre, Brazil.
- Gonzatto, M. P., Schwarz, S. F., Madail, J. C. M., Campos, A. D., Nava, D. E., and Ueno, B. 2010. Sistemas agroflorestais. In: Oliveira, R. P., Scivittaro, W. B., Schroder, E. C., *et al.* (eds.). *Produção orgânica de citros no Rio Grande do Sul*. Embrapa Clima Temperado, Pelotas, Brazil, pp. 219-232.
- Haile, S. G., Nair, P. K., and Nair, V. D. 2008. Carbon storage of different soil-size fractions in Florida silvopastoral systems. *Journal of Environmental Quality*, 37: 1789-1797.
- Harvey, C. A., and Villalobos, J. A. G. 2007. Agroforestry systems conserve species-rich but modified assemblages of tropical birds and bats. *Biodiversity and Conservation*, 16: 2257-2292.
- Hawke, M. F., and O'Connor, M. B. 1993. Soil pH and nutrient levels at Tikitere agroforestry research area. *New Zealand Journal of Forestry Science*, 23: 40-48.
- IBGE – Instituto Brasileiro de Geografia e Estatística. 2009. Censo Agropecuário 2006 – Brasil, Regiões e Unidades da Federação. IBGE, Rio de Janeiro, Brasil.
- IBGE - Instituto Brasileiro de Geografia e Estatística. 2016. SIDRA - Sistema IBGE de Recuperação Automática. IBGE, Brasília, Brasil, 2012. Available on: <http://www.sidra.ibge.gov.br> Accessed in March 28, 2016.
- Ichinose, K., Hoa, N. V., Bang, D. V., Tuan, D. H., and Dien, L. Q. 2012. Limited efficacy of guava interplanting on citrus greening disease: Effectiveness of protection against disease invasion breaks down after one year. *Crop Protection*, 34: 119-126.
- Jifon, J. L., and Syvertsen, J. P. 2003. Moderate shade can increase net gas exchange and reduce photoinhibition in citrus leaves. *Tree Physiology*, 23: 119-127.
- Kim, D. G., Kirschbaum, M. U., and Beedy, T. L. 2016. Carbon sequestration and net emissions of CH₄ and N₂O under agroforestry: Synthesizing available data and suggestions for future studies. *Agriculture, Ecosystems & Environment*, 226: 65-78.
- Knapp, J. R., Laur, G. L., Vadas, P. A., Weiss, W. P., and Tricarico, J. M. 2014. Enteric methane in dairy cattle production: quantifying the opportunities and impact of reducing emissions. *Journal of Dairy Science*, 97: 3231-3261.
- Kremen, C., and Miles, A. 2012. Ecosystem services in biologically diversified versus conventional farming systems: benefits, externalities, and trade-offs. *Ecology and Society*, 17: art40.
- Kriedemann, P. E. 1968. Some photosynthetic characteristics of citrus leaves. *Australian Journal of Biological Sciences*, 21: 895-906.
- Kumar, B. M. 2008. Litter dynamics in plantation and agroforestry systems of the tropics – a review of observations and methods. In: Batish, D. R., Kohli, R. K., Jose, S., and Singh, H. P. (eds.) *Ecological basis of agroforestry*. CRC Press, New York, U. S. A., pp. 181-216.
- Lagier, J. 2008. La aventura de la Yerba Mate: más de cuatro siglos de historia. [s.e.], Buenos Aires, Argentine.
- Leite, P. F. 1995. As diferentes unidades fitoecológicas do sul do Brasil – proposta de classificação. *Cadernos de Geociências*, 15: 73-164.

- Linhares, T. 1969. História econômica do mate. Livraria José Olympio Editora, Rio de Janeiro, Brazil.
- Liu, C., and Liu, Y. 2012. Impacts of shading in field on micro-environmental factors around plants and quality of pineapple fruits. *Journal of Food Agriculture & Environment*, 10: 741-745.
- Lorenz, K., and Lal, R. 2014. Soil organic carbon sequestration in agroforestry systems. A review. *Agronomy for Sustainable Development*, 34: 443-454.
- Maas, B., Clough, Y., and Tschardtke, T. 2013. Bats and birds increase crop yield in tropical agroforestry landscapes. *Ecology Letters*, 16: 1480-1487.
- Magalhães, J. V., Costa, N. L., Pereira, R. G. de A., Townsend, C. R. 2001. Desempenho produtivo e reações fisiológicas de ovinos deslanados mantidos sob seringal (*Hevea brasiliensis*). *Revista Científica de Produção Animal*, 3: 77-82.
- Maia, S. M. F., Xavier, F. A. S., Oliveira, T. S., Mendonça, E. S., and Araújo Filho, J. A. 2007. Organic carbon pools in a Luvisol under agroforestry and conventional farming systems in the semi-arid region of Ceará, Brazil. *Agroforestry Systems*, 71: 127-138.
- Makumba, W., Akinnifesi, F. K., Janssen, B., and Oenema, O. 2007. Long-term impact of a gliricidia-maize intercropping system on carbon sequestration in southern Malawi. *Agriculture, Ecosystems & Environment*, 118: 237-243.
- Martin, F. S., and Van Noordwijk, M. 2009. Trade-offs analysis for possible timber-based agroforestry scenarios using native trees in the Philippines. *Agroforestry Systems*, 76: 555-567.
- McGinty, M. M., Swisher, M. E., and Alavalapati, J. 2008. Agroforestry adoption and maintenance: self-efficacy, attitudes and socio-economic factors. *Agroforestry Systems*, 73: 99-108.
- McNeely, J. A., and Schroth, G. 2006. Agroforestry and biodiversity conservation—traditional practices, present dynamics, and lessons for the future. *Biodiversity and Conservation*, 15: 549-554.
- Meijer, S. S., Sileshi, G. W., Catacutan, D., and Nieuwenhuis, M. 2016. Agroforestry and deforestation in Malawi: inter-linkages between attitudes, beliefs and behaviours. *Agroforestry Systems*, 90: 645-658.
- Mercer, D. E. 2004. Adoption of agroforestry innovations in the tropics: a review. *Agroforestry Systems*, 61: 311-328.
- Miccolis, A., Vivan, J. L., Gonçalves, A. L., Meier, M., and Porro, R. 2011. Políticas públicas e Sistemas Agroflorestais: lições aprendidas a partir de cinco estudos de caso no Brasil. In: Miccolis, A. and Porro, R. (eds.) *Políticas Públicas para o Desenvolvimento Agroflorestal no Brasil*. ICRAF, Belém do Pará, pp. 1-24.
- Mognon, F., Dallagnol, F. S., Sanquetta, C. R., Corte, A. P. D., and Barreto, T. G. 2013. Uma década de dinâmica da fixação de carbono na biomassa arbórea em Floresta Ombrófila Mista no sul do Paraná. *Floresta*, 43: 153-164.
- Molina, R. D. O., Nunes, W. M. D. C., Gil, L. G., Rinaldi, D. A. M. D. F., Croce Filho, J., and Carvalho, R. C. Z. D. 2014. First Report of Citrus *Aleurocanthus woglumi* Ashby (Hemiptera: Aleyrodidae) in the State of Paraná, Brazil. *Brazilian Archives of Biology and Technology*, 57: 472-475.
- Mpairwe, D. R., Sabiiti, E. N., and Mugerwa, J. S. 1998. Effect of dried *Gliricidia sepium* leaf supplement on feed intake, digestibility and nitrogen retention in sheep fed dried KW4 elephant grass (*Pennisetum purpureum*) ad libitum. *Agroforestry Systems*, 41: 139-150.
- Muñoz, C., Zagal, E., and Ovalle, C. 2007. Influence of trees on soil organic matter in Mediterranean agroforestry systems: an example from the ‘Espinal’ of central Chile. *European Journal of Soil Science*, 58: 728-735.
- Murgueitio, E., Calle, Z., and Uribe, F. 2011. Native trees and shrubs for the productive rehabilitation of tropical cattle ranching lands. *Forest Ecology & Management*, 261: 1654–1663.
- Muriel, R., Sandra, B., Muñoz, G., and Restrepo, D. 2014. Parasitoidismo de dos especies de mariposas en dos sistemas de producción de café. *Revista Colombiana de Entomología*, 40: 251-258.
- Mutuo, P. K., Cadisch, G., Albrecht, A., Palm, C. A., and Verchot, L. 2005. Potential of agroforestry for carbon sequestration and mitigation of greenhouse gas emissions from soils in the tropics. *Nutrient Cycling in Agroecosystems*, 71: 43–54.

- Nair, P. K. R., and Garrity, D. P. 2012. Agroforestry - the future of global land use. Springer, Dordrecht, Netherlands.
- Nair, P. K. R., Mohan Kumar, B., and Nair, V. D. 2009. Agroforestry as a strategy for carbon sequestration. *Journal of Plant Nutrition and Soil Science*, 172: 10-23.
- Nath, C. D., Schroth, G., and Burslem, D. F. R. P. 2016. Why do farmers plant more exotic than native trees? A case study from the Western Ghats, India. *Agriculture, Ecosystems & Environment*, 230: 315-328.
- Ngwabie, N. M., Jeppsson, K. H., Gustafsson, G., and Nimmermark, S. 2011. Effects of animal activity and air temperature on methane and ammonia emissions from a naturally ventilated building for dairy cows. *Atmospheric Environment*, 45: 6760-6768.
- Norgrove, L. 1998. Musa in multistrata systems: focus on shade. *Infomusa*, 7: 17-22.
- Norgrove, L., and Hauser, S. 2013. Black leaf streak disease and plantain fruit characteristics as affected by tree density and biomass management in a tropical agroforestry system. *Agroforestry Systems*, 87: 349-354.
- Nygren P., Fernández, M. P., Harmand, J. M., and Leblanc, H. A. 2012. Symbiotic dinitrogen fixation by trees: an underestimated resource in agroforestry systems? *Nutrient Cycling in Agroecosystems*, 94: 123-160.
- Paciullo, D. S. C., Castro, C. R. T., Gomide, C. A. D. M., Maurício, R. M., Pires, M. D. F. Á., Müller, M. D., and Xavier, D. F. 2011. Performance of dairy heifers in a silvopastoral system. *Livestock Science*, 141: 166-172.
- Pagliosa, C. M., Pereira, S. M., Vieira, M. A., Costa, L. A., Teixeira, E., Amboni, R. D., and Amante, E. R. 2009. Bitterness in yerba mate (*Ilex paraguariensis*) leaves. *Journal of Sensory Studies*, 24: 415-426.
- Pattanayak, S. K., Mercer, D. E., Sills, E., and Yang, J. C. 2003. Taking stock of agroforestry adoption studies. *Agroforestry Systems*, 57: 173-186.
- Perez, J. O., Nova, F. A., Portillo, B. A., Ortega, O. A. C., and Hernandez, S. R. 2013. Use of three fodder trees in the feeding of goats in the subhumid tropics in Mexico. *Tropical Animal Health and Production*, 45: 821-828.
- Philpott, S. M., and Armbrrecht, I. 2006. Biodiversity in tropical agroforests and the ecological role of ants and ant diversity in predatory function. *Ecological Entomology*, 31: 369-377.
- Pillar, V. D. P., and Vélez, E. 2010. Extinção dos Campos Sulinos em Unidades de Conservação: um fenômeno natural ou um problema ético? *Natureza & Conservação*, 8: 84-86.
- Pompéia, S. 2005. Recuperação da vegetação da Serra do Mar em áreas afetadas pela poluição atmosférica de Cubatão: uma análise histórica. In: Galvão, A. P. M., Porfírio-Da-Silva, V. (eds.). *Restauração Florestal – Fundamentos e Estudos de Caso*. Embrapa Florestas, Colombo, Brazil, pp.119-143.
- Pontes, L. D. S., Barro, R. S., Camargo, E. F., da Silva, V. P., Cezimbra, I., Berndt, A., Bayer, C., and Carvalho, P. C. D. F. 2014. Methane emissions from ruminants in integrated crop-livestock systems. *Tropical Grasslands*, 2: 124-126.
- Primieri, S., Dalla Costa, M., Stroschein, M. R. D., Stocco, P., Santos, J. C. P., and Antunes, P. M. 2016. Variability in symbiotic effectiveness of N₂ fixing bacteria in *Mimosa scabrella*. *Applied Soil Ecology*, 102: 19-25.
- Pumariño, L., Sileshi, G. W., Gripenberg, S., Kaartinen, R., Barrios, E., Muchane, M. N., Midega, C. and Jonsson, M. 2015. Effects of agroforestry on pest, disease and weed control: a meta-analysis. *Basic and Applied Ecology*, 16: 573-582.
- Rapidel, B., Allinne, C., and Cerdán, C. 2015. Efectos ecológicos y productivos del asocio de árboles de sombra con café em sistemas agroforestales. In: Montagnini, F., Somarriba, E., Murgueitio, E. (eds.) *Sistemas Agroforestales - Funciones Productivas, Socioeconómicas y Ambientales*. CATIE, Turrialba, Costa Rica, pp. 5-20.
- Rivera, L. F., Armbrrecht, I., and Calle, Z. 2013. Silvopastoral systems and ant diversity conservation in a cattle-dominated landscape of the Colombian Andes. *Agriculture, Ecosystems & Environment*, 181: 188-194.
- Rodrigues, E. R., Júnior, L. C., Moscolgiato, A. V., and Beltrame, T. P. 2008. O uso do sistema agroflorestal Taungya na restauração de reservas legais: indicadores econômicos. *Floresta*, 38: 517-525.
- Rodríguez, J. A., and García, J. C. C. 2009. Erosión y escorrentía: indicadores de respuesta temprana del suelo a distintas coberturas en la zona cafetalera de Colombia. *Recursos Naturales y Ambiente*, 58: 25-31.
- Rowlings, D. W., Grace, P. R., Kiese, R., and Weier, K. L. 2012. Environmental factors controlling

- temporal and spatial variability in the soil-atmosphere exchange of CO₂, CH₄ and N₂O from an Australian subtropical rainforest. *Global Change Biology*, 18: 726-738.
- Saibro, J. C., Castilhos, Z. M. S., Silva, J. L. S., Varella, A. C., Lucas, N. M., and Savian, J. F. 2009. A integração da silvicultura com pastagens e pecuária no Rio Grande do Sul. In: Pillar, V. D. P., Müller S. C., Castilhos Z. M. D. S., and Jacques, A. V. Á. (eds.), *Campos Sulinos-conservação e uso sustentável da biodiversidade*. Ministério do Meio Ambiente-MMA, Brasília, Brazil, p. 260-265.
- Santos, M. V. F. D., Lira, M. D. A., Junior, D., Batista, J. C., Guim, A., Mello, A. C. L. D., and Cunha, M. V. D. 2010. Potential of Caatinga forage plants in ruminant feeding. *Revista Brasileira de Zootecnia*, 39: 204-215.
- Schmidt, M. W., Torn, M. S., Abiven, S., Dittmar, T., Guggenberger, G., Janssens, I. A., Kleber, M., Kögel-Knabner, I., Lehmann, J., Manning, D. A. C., Nannipieri, P., Rasse, D. P., Weiner, S. and Trumbore, S. E. 2011. Persistence of soil organic matter as an ecosystem property. *Nature*, 478: 49-56.
- Schroth, G., Krauss, U., Gasparotto, L., Aguilar, J. D., and Vohland, K. 2000. Pests and diseases in agroforestry systems of the humid tropics. *Agroforestry systems*, 50: 199-241.
- Schwiderke, D. K., Cezar, R. M., Vezzani, F. M., Froufe, L. C. M., and Seone, C. E. S. 2012. Atributos químicos do solo em sistemas agroflorestais multiestrata sucessional e em áreas de regeneração natural. In: *Proceedings of Congresso Florestal Paranaense; 2012 Sep 10-14; Curitiba, Brazil*. Curitiba: EMBRAPA, 2002. Available in <<https://ainfo.cnptia.embrapa.br/digital/bitstream/item/66949/1/LuisF-CFP-AtributosQuimicos-1.pdf>> Accessed in May 05, 2017.
- Sharma, G., Sharma, R., and Sharma, E. 2009. Impact of stand age on soil C, N and P dynamics in a 40-year chronosequence of alder-cardamom agroforestry stands of the Sikkim Himalaya. *Pedobiologia*, 52: 401-414.
- Signor, P. 2013. Biomassa comercial de *Ilex paraguariensis* St.-Hil. e sua relação com variáveis ambientais em floresta com araucária, Paraná. Master Dissertation, UNICENTRO, Irati, Brazil.
- Silva, A. G. D., Boiça Junior, A. L., Farias, P. R. S., and Barbosa, J. C. 2011a. Infestation of citrus blackfly in citrus (*Aleurocanthus woglumi* Ashby) orchards conventional and agroforestry. *Revista Brasileira de Fruticultura*, 33: 53-60.
- Silva, G. L., Lima, H. V., and Campanha, M. M. 2011b. Soil physical quality of Luvisols under agroforestry, natural vegetation and conventional crop management systems in the Brazilian semi-arid region. *Geoderma*, 167: 61-70.
- Somarriba, E., and Kass, D. 2001. Estimates of above-ground biomass and nutrient accumulation in Mimosa scabrella fallows in southern Brazil. *Agroforestry Systems*, 51: 57-84.
- Souza, F. M., and Batista, J. L. F. 2004. Restoration of seasonal semideciduous forests in Brazil: influence of age and restoration design on forest structure. *Forest Ecology & Management*, 191: 185-200.
- Souza, H. N., Cardoso, I. M., Sá Mendonça, E., Carvalho, A. F., de Oliveira, G. B., Gjorup, D. F., and Bonfim, V. R. 2012. Learning by doing: a participatory methodology for systematization of experiments with agroforestry systems, with an example of its application. *Agroforestry Systems*, 85: 247-262.
- Spiertz, J. H. J. 2010. Nitrogen, sustainable agriculture and food security. A review. *Agronomy for Sustainable Development*, 30: 43-55.
- Steffan-Dewenter, I., Kessler, M., Barkmann, J., Bos, M. M., Buchori, D., Erasmi, S., Faust, H., Gerold, G., Glenk, K., Gradstein, S. R., Guhardja, E., Harteveld, M., Hertel, D., Höhn, P., Kappas, M., Köhler, S., Leuschner, C., Maertens, M., Marggraf, R., Migge-Kleian, S., Moge, J., Pitopang, R., Schaefer, M., Schwarze, S., Sporn, S. G., Steingrebe, A., Tjitrosoedirdjo, S. S., Tjitrosoemito, S., Twele, A., Weber, R., Woltmann, L., Zeller, M., and Tschardtke, T. 2007. Tradeoffs between income, biodiversity, and ecosystem functioning during tropical rainforest conversion and agroforestry intensification. *Proceedings of National American Society*, 104: 4973-4978.
- Streit, N. M., Hecktheuer, L. H. R., Do Canto, M. W., Mallmann, C. A., Streck, L., Parodi, T. V., and Canterle, L. P. 2007. Relation among taste-related compounds (phenolics and caffeine) and sensory profile of erva-mate (*Ilex paraguariensis*). *Food Chemistry*, 102: 560-564.

- Syvertsen, J. P., Goñi, C., and Otero, A. 2003. Fruit load and canopy shading affect leaf characteristics and net gas exchange of 'Spring' navel orange trees. *Tree Physiology*, 23: 899-906.
- Tamang, B., Andreu, M. G., and Rockwood, D. L. 2010. Microclimate patterns on the leeward side of single-row tree windbreaks during different weather conditions in Florida farms: implications for improved crop production. *Agroforestry Systems*, 79: 111-122.
- Tonin, J., Poester, G. C., Andriolli, E. M., Pelissari, J. C., Giraldo, P. E. C., and Ignace, A. D. 2017. Cadeia solidária das frutas nativas: algumas reflexões a respeito da segurança alimentar e nutricional. *Revista Brasileira de Extensão Universitária*, 8: 49-56.
- Traore, K., Ganry, F., Oliver, R., and Gigou, J. 2004. Litter production and soil fertility in a *Vitellaria paradoxa* parkland in a Catena in Southern Mali. *Arid Land Research and Management*, 18: 359-368.
- Tscharntke, T., Sekercioglu, C. H., Dietsch, T. V., Sodhi, N. S., Hoehn, P., and Tylianakis, J. M. 2008. Landscape constraints on functional diversity of birds and insects in tropical agroecosystems. *Ecology*, 89: 944-951.
- Tscharntke, T., Clough, Y., Bhagwat, S. A., Buchori, D., Faust, H., Hertel, D., Hölscher, D., Jührbandt, J., Kessler, M., Perfecto, I., Scherber, C., Schroth, G., Veldkamp, E., and Wanger, T. C. 2011. Multifunctional shade-tree management in tropical agroforestry landscapes—a review. *Journal of Applied Ecology*, 48: 619-629.
- Udawatta, R. P., and Jose, S. 2012. Agroforestry strategies to sequester carbon in temperate North America. *Agroforestry Systems*, 86: 225-242.
- Vallejo, V. E., Roldan, F., and Dick, R. P. 2010. Soil enzymatic activities and microbial biomass in an integrated agroforestry chronosequence compared to monoculture and a native forest of Colombia. *Biology and Fertility of Soils*, 46: 577-587.
- Van der Werf, W., Keesman, K., Burgess, P. J., Graves, A., Pilbeam, D., Incoll, L. D., Metselaar, K., Mayus, M., Stappers, R., van Keulen, H., and Palma, J. 2007. Yield-SAFE: a parameter-sparse process-based dynamic model for predicting resource capture, growth and production in agroforestry systems. *Ecological Engineering*, 29: 419-433.
- Verchot, L. V. J., Mackensen, J., Kandji, S., van Noordwijk, M., Tomich, T., Ong, C., Albrecht, A., Bantilan, C., Anupama, K. V., and Palm, C. 2005. Opportunities for linking adaptation to climate change. In: Robledo, C., Kanninen, M., and Pedroni, L. (eds). *Tropical forest and adaptation to climate changes: search of synergies*. Center for Forestry Research (CIFOR), Bogor, Indonesia, pp. 103-121.
- Verchot, L. V. J., Van Noordwijk, M., and Kandji, S. 2007. Climate change: Linking adaptation and mitigation through agroforestry. *Mitigation and Adaptation Strategies for Global Change*, 12: 901-918.
- Vieira, D. L., Holl, K. D., and Peneireiro, F. M. 2009. Agro-successional restoration as a strategy to facilitate tropical forest recovery. *Restoration Ecology*, 17: 451-459.
- Wen-Jie, W., Ling, Q., Yuan-Gang, Z., Dong-Xue, S., Jing, A., Hong-Yan, W., Guan-Yu, Z., Wei, S., and Xi-Quan, C. 2011. Changes in soil organic carbon, nitrogen, pH and bulk density with the development of larch (*Larix gmelinii*) plantations in China. *Global Change Biology*, 17: 2657-2676.
- Wheaton, T. A., Castle, W. S., and Tucker, D. P. H. 1978. Higher density plantings for Florida citrus - concepts. *Proceedings of Florida State Horticultural Society*, 91: 27-33.
- Wilson, M. H., and Lovell, S. T. 2016. Agroforestry - the next step in sustainable and resilient agriculture. *Sustainability*, 8: 574.
- Yamamoto, W., Ap Dewi, I., and Ibrahim, M. 2007. Effects of silvopastoral areas on milk production at dual-purpose cattle farms at the semi-humid old agricultural frontier in central Nicaragua. *Agroforestry Systems*, 94: 368-375.
- Yu, C. M. 2004. Seqüestro florestal de carbono no Brasil - dimensões políticas, socioeconômicas e ecológicas. *Annablume*, São Paulo, Brazil.
- Zanella, J. B., and Coelho, G. C. 2014. The effect of plant biomass from pruning of trees on maize (*Zea mays* L.) and weeds. *Revista Brasileira de Agroecologia*, 9: 54-62.
- Ziller, S., Carpanezzi, O. T. B., and Carpanezzi, A. A. 2010. Guia técnico para recuperação ambiental de áreas protegidas na região do noroeste do estado do Paraná. DIBAP, Curitiba, Brazil.